

1 © <2021>. This manuscript version is made available under the CC-BY-NC-ND 4.0  
2 license <http://creativecommons.org/licenses/by-nc-nd/4.0/>

3 This document is the Accepted Manuscript version of a Published Work that appeared in final form in  
4 [Developmental and Comparative Immunology]. To access the final edited and published work see  
5 [[http:// dx.doi.org/10.1016/j.dd.2017.06.002](http://dx.doi.org/10.1016/j.dd.2017.06.002)].”

6

7

## 8 **17 $\alpha$ -Ethinylestradiol alters the peritoneal immune response of gilthead seabream**

9

10 NE Gómez González<sup>a</sup>, I Cabas<sup>a</sup>, MC Rodenas<sup>a</sup>, M Arizcun<sup>b</sup>, V Mulero<sup>a</sup>, A García Ayala<sup>a\*</sup>

11 <sup>a</sup>Department of Cell Biology and Histology, Faculty of Biology, Regional Campus of  
12 International Excellence “Campus Mare Nostrum”, University of Murcia, IMIB-Arrixaca,  
13 30100 Murcia, Spain

14 <sup>b</sup>Centro Oceanográfico de Murcia, Instituto Español de Oceanografía (IEO), Carretera de la  
15 Azohía s/n, Puerto de Mazarrón, 30860 Murcia, Spain

16

17 \*To whom correspondence should be addressed at Department of Cell Biology and Histology,  
18 Faculty of Biology, University of Murcia, Espinardo, 30100. Murcia. Spain. Phone: +34 868  
19 884968. Fax: +34 868 883963. Email: [agayala@um.es](mailto:agayala@um.es)

20 **Key words:** 17 $\alpha$ -ethinylestradiol, peritoneal exudate, immune response, gilthead seabream

21

## 22 **Abstract**

23  $17\alpha$ -Ethinylestradiol (EE<sub>2</sub>), a synthetic estrogen used in most oral contraceptives pills and  
24 hormone replacement therapies, is found in many water bodies, where it can modulate the fish  
25 immune response. EE<sub>2</sub> acts as an endocrine disruptor in gilthead seabream, *Sparus aurata* L., a  
26 marine teleost fish of great economic value in Mediterranean aquaculture, as it induces hepatic  
27 vitellogenin gene (*vtg*) expression. Moreover, EE<sub>2</sub> also alters the capacity of gilthead seabream  
28 to appropriately respond to infection although it does not behave as an immunosuppressor.  
29 Nevertheless, these previous studies have mainly focused on the head kidney leukocytes and  
30 no information exists on peritoneal leukocytes, including mast cells. In the present work,  
31 juvenile gilthead seabream fish were fed a pellet diet supplemented with EE<sub>2</sub> for 76 days and  
32 intraperitoneally injected with hemocyanin plus imject alum adjuvant at the end of EE<sub>2</sub>  
33 treatment and 92 days later, and the peritoneal immune response was analyzed. EE<sub>2</sub>  
34 supplementation induced *vtg* expression but returned to basal levels by 3 months post-treatment.  
35 Interestingly, gilthead seabream peritoneal leukocytes express the genes encoding for the  
36 nuclear estrogen receptor  $\alpha$  and the G protein-coupled estrogen receptor 1 and the dietary intake  
37 of EE<sub>2</sub> induced these expression. Moreover, EE<sub>2</sub> induced an inflammatory response in the  
38 peritoneal cavity in unvaccinated fish, which was largely maintained for several months after  
39 the cessation of the treatment. However, the impact of EE<sub>2</sub> in vaccinated fish was rather minor  
40 and transient. Taken together, the study provides fresh information about endocrine immune  
41 disruption, focusing on peritoneal leukocytes.

## 42 1. Introduction

43 It is well known that estrogens are key modulators of the immune system (Straub,  
44 2007), including that of fish, due to the presence of nuclear estrogen receptors (ERs) in  
45 piscine leukocytes (Cabas et al., 2013a; Liu et al., 2009; Lynn et al., 2008; Shved et al.,  
46 2009; Slater et al., 1995; Todo et al., 1999). Moreover, the identification of a G protein-  
47 coupled estrogen receptor 1 (GPER1), a membrane estrogen receptor that binds 17 $\beta$ -  
48 estradiol (E<sub>2</sub>) (Revankar et al., 2005; Thomas et al., 2005) and other estrogens (Prossnitz  
49 and Hathaway, 2015), and its identification in immune cells of both mammals and fish  
50 (Blasko et al., 2009; Brunsing et al., 2013; Cabas et al., 2013a), has opened up the  
51 possibility of exploring additional estrogen-mediated effects.

52 Anthropogenic activity produces the release of a wide variety of chemical compounds into  
53 the environment, most of which affect in some way the species living there. Among these  
54 compounds, endocrine disruptor chemicals (EDCs) are a group of substances able to  
55 mimic and/or antagonize the effects of endogenous endocrine hormones, and disrupt the  
56 synthesis and metabolism of endogenous hormones and the synthesis of hormone  
57 receptors (Aris et al., 2014). EDCs alter fish reproduction, but also growth,  
58 osmoregulation and immunity and induce stress (Filby et al., 2007).

59 17 $\alpha$ -Ethinylestradiol (EE<sub>2</sub>), considered a xeno-estrogen EDC, is a synthetic estrogen  
60 and the main active component of contraceptive pills, which are also used in hormonal  
61 therapy. EE<sub>2</sub> can be found not only in sewage treatment works (Baronti et al., 2000;  
62 Desbrow et al., 1998; Quinn et al., 2004; Vethaak et al., 2005). Due to its capacities to be  
63 bio-accumulated and to alter endocrine homeostasis, much research effort has focused on  
64 its effects on animals living in aquatic environments (Aris et al., 2014). It has been found  
65 that EE<sub>2</sub> can modify the expression of ERs and the expression of the genes classically  
66 regulated by these receptors (Milla et al., 2011), such as the vitellogenin (vtg) gene  
67 (Kausch et al., 2008), which is widely accepted to be a biomarker of endocrine disruption  
68 by xeno-estrogens (Sumpter and Jobling, 1995).

69 The gilthead seabream (*Sparus aurata* L.), a marine protandrous teleost fish of great  
70 commercial value in the Mediterranean Sea and in aquaculture, has been widely used to  
71 study the effect of EE<sub>2</sub> on the reproductive (Cabas et al., 2011) and immune systems at  
72 different times of its reproductive cycle (Cabas et al., 2013b; Cabas et al., 2012; Rodenas  
73 et al., 2015; Rodenas et al., 2016). Thus, EE<sub>2</sub> provokes an estrogenic response, inducing

74 hepatic *vtg* gene expression (Cabas et al., 2013b; Cabas et al., 2012; Rodenas et al., 2015;  
75 Rodenas et al., 2016), and alters the expression of the gene encoding for the ER $\alpha$  of head  
76 kidney leukocytes (Cabas et al., 2012). EE<sub>2</sub> might alter the capacity of fish to  
77 appropriately respond to infection although it does not behave as an immunosuppressor,  
78 as assayed *in vivo* and *in vitro* using head kidney leukocytes (Cabas et al., 2012; Rodenas  
79 et al., 2015). Interestingly, the estrogenic response of EE<sub>2</sub> disappears when exposure to it  
80 ceases, although the adaptive immune response remains altered to some extent (Rodenas  
81 et al., 2015; Rodenas et al., 2016). To increase our knowledge of the effect of EE<sub>2</sub> on the  
82 immune response of gilthead seabream, we focus our study on its effects on peritoneal  
83 cells. The peritoneal exudate of healthy gilthead seabream, as occurs in sea bass, consists  
84 of granulocytes, macrophages, lymphocytes and mast cells (Gómez-González et al., 2014;  
85 Meseguer et al., 1993) and its cellular composition may vary depending on the extent of  
86 tissue damage or injury. Although some studies have revealed the effect of estrogens,  
87 both natural and synthetic, on peritoneal leukocytes, there is a noticeable lack of  
88 information about the effects of xenoestrogens on piscine peritoneal leukocytes, and, to  
89 the best of our knowledge, no information exists concerning how they may be modulated  
90 by EE<sub>2</sub>.

91 The aim of this work was, therefore, to evaluate the impact of dietary administration  
92 of EE<sub>2</sub> in leukocyte activation (respiratory burst and cytokine gene expression), and mast  
93 cell and acidophilic granulocyte abundance in the peritoneal exudate of fish upon  
94 intraperitoneal (i.p.) vaccination

## 95 **2. Materials and Methods**

### 96 **2.1. Animals and experimental design**

97 Healthy specimens of gilthead seabream *Sparus aurata* L. (Actinopterygii,  
98 Perciformes, Sparidae) were reared at the Centro Oceanográfico de Murcia (Instituto  
99 Español de Oceanografía, Mazarrón, Murcia, Spain), where they were kept in running  
100 seawater aquaria (dissolved oxygen 6 ppm, flow rate 20% aquarium vol/h) with a natural  
101 temperature and photoperiod, and fed three times per day with a commercial pellet diet  
102 (44% protein, 22% lipids; Skretting) at a feeding rate of 1.5% of fish biomass.  
103 Environmental parameters, mortality and food intake, as well as behaviour, were recorded  
104 daily. The experiments described were approved by the Consejería de Agua, Agricultura  
105 y Medio Ambiente of the Región de Murcia (approval number A13160507).

106 *In vivo* EE<sub>2</sub> treatment was carried out with juvenile gilthead seabream specimens (n  
107 = 400) with a body weight of 26.6 ± 4.2 g in 170 L aquaria. Briefly, EE<sub>2</sub> (5 µg/g food,  
108 98% purity; Sigma) was incorporated in the commercial food using the ethanol  
109 evaporation method (0.3 L ethanol/kg of food), as described previously (Shved et al.,  
110 2007). The specimens were fed three times a day *ad libitum* with the pellet diet  
111 supplemented with EE<sub>2</sub> (treated fish) or the pellet diet prepared with ethanol but without  
112 EE<sub>2</sub> (untreated fish) for 76 days (days of treatment, dot) after which they were fed with a  
113 commercial feed for a further 142 days. In order to evaluate the effect of EE<sub>2</sub> on an  
114 induced immune response, the specimens were i.p. injected with hemocyanin (45 µg/fish;  
115 Sigma-Aldrich) plus alum adjuvant (3.6 µg/fish; Thermo Scientific) (vaccinated fish) or  
116 phosphate buffered saline (PBS) (unvaccinated fish) 1 day after treatment (dat) (priming)  
117 and 92 dat (booster) (Supplementary Fig. 1).

118 Samples of liver and peritoneal exudate leukocytes were collected 1, 93 and 142 dat  
119 (1 day post-priming, dpp and 1 and 50 days post-booster, dpb, respectively)  
120 (Supplementary Fig. 1). Specimens (n=6 fish/treatment/time of sampling) were fasted for  
121 24 h before each sampling. They were tranquilized by 8 µl/l of clove oil and immediately  
122 anesthetized using 40 µl/l of clove oil and weighed. Two-three mL of sRPMI culture  
123 medium [RPMI culture medium (Sigma-Aldrich) with 2 mM glutamine (Sigma-Aldrich),  
124 100 i.u./ml penicillin and 100 µg/ml streptomycin (Sigma-Aldrich) adjusted with 0.35%  
125 NaCl to seabream serum osmolarity (353.33 mOs)] were i.p. injected at 1 dpp, 1 dpb and  
126 50 dpb. Fish abdomens were massaged for 10 min to dislodge tissue-attached cells into  
127 the sRPMI culture medium. Then, incisions were made below of the lateral fin to access  
128 the peritoneum and the peritoneal exudates were aspirated and collected into 15 mL  
129 Falcon tubes, after which the specimens were decapitated and the liver samples were keep  
130 on RNAlater (Sigma-Aldrich) and on ice until analysis.

## 131 **2.2. Peritoneal exudate cell suspensions**

132 Peritoneal exudates (approximately 2-3 mL) were passed through 70 µm cell  
133 strainers in order to remove large clumps of cells. Then, the cells were centrifuged at 600  
134 x g for 5 min and resuspended in 5 mL of sRPMI. The total number of cells in peritoneal  
135 exudate was counted using an automatic cell counter (Bio-Rad).

136 **2.3. Determination of the percentage of acidophilic granulocytes and mast cells in**  
137 **peritoneal exudate cell suspensions**

138 The percentage of acidophilic granulocytes and mast cells was analyzed using 0.1 x  
139 10<sup>6</sup> peritoneal exudate cell suspensions. Cells were incubated on ice with a specific  
140 monoclonal antibody (mAb) against gilthead seabream acidophilic granulocytes (G7,  
141 1:100) (Sepulcre et al., 2002), to measure the percentage of acidophilic granulocytes, or  
142 with a mAb against seabream mast cells (GB10, 1/1) (Gómez-González et al., 2016), to  
143 determine the percentage of mast cells in 100 µL of FACS buffer (PBS, 0.35% NaCl, 2%  
144 fetal bovine serum (FBS, Sigma-Aldrich) and 0.05% sodium azide (Sigma-Aldrich).  
145 After 30 min, cells were centrifuged at 600 x g for 5 min, the supernatant was removed  
146 by aspiration and the cells were washed twice with FACS buffer. Then, cells were labelled  
147 with the polyclonal antibody Alexa Fluor 488 F(ab')<sub>2</sub> fragment (1:500) of goat anti-mouse  
148 IgG (H+L) (ThermoFisher Scientific) for 30 min at 4 °C. The supernatant was removed  
149 and cells were washed, as previously described. Fluorescence intensity was measured by  
150 flow cytometry using a FACSCalibur flow cytometer (BD Biosciences) and data were  
151 analyzed with FlowJo v10.0.4 software.

152 **2.4. ROS production assay**

153 ROS production was measured using luminol-dependent chemiluminescence  
154 produced by 0.5 x 10<sup>6</sup> peritoneal exudate cell suspensions. Cells were incubated at 21 °C  
155 and 5% CO<sub>2</sub> during 20 min. Then, the ROS production was determined by adding 100  
156 µM luminol (Sigma-Aldrich) and 1 µg/mL phorbolmyristate acetate (PMA) (Sigma-  
157 Aldrich) in Hanks' Balanced Salt Solution (Sigma-Aldrich) to the cell culture.  
158 Chemiluminescence was recorded every 127 s for 1 h using a FLUOstartluminometer  
159 (BGM, LabTechnologies). Three experimental replicates were performed for all samples  
160 (6). The values reported are the average of the replicate samples, expressed as maximum  
161 of the reaction curve from 127 to 1016 s, from which the background (only culture  
162 medium) was subtracted.

163 **2.5. Gene expression analysis**

164 Liver samples were defrosted and RNAlater removed. Total RNA was extracted from  
165 liver and peritoneal exudate cell suspensions of 1, 93 and 142 dat (1 dpp, 1dpb and 50  
166 dpb, respectively) with TRIzol Reagent (ThermoFisher Scientific) as indicated by the  
167 manufacturer's instructions, and the RNA concentration was quantified by

168 spectrophotometry (NanoDrop, ND-1000). RNA was treated with DNase I (amplification  
169 grade, 1 unit/ $\mu$ g RNA, Invitrogen) to remove genomic DNA traces that could interfere  
170 with the PCR reactions. SuperScript III RNase H- Reverse Transcriptase (Invitrogen) was  
171 used to synthesize first strand cDNA with oligo (dT)<sub>18</sub> primer (Invitrogen) from 1  $\mu$ g of  
172 total RNA, at 50 °C for 50 min. The quality of cDNA was analyzed by performing a semi-  
173 quantitative PCR of  $\beta$ -actin (*actb*) housekeeping gene using an Eppendorf Mastercycler  
174 Gradient Instrument (Eppendorf). The reaction mixture was incubated for 2 min at 94 °C,  
175 followed by 30 cycles of 45 s at 94 °C, 45 s at 55 °C (the specific annealing temperature),  
176 and 45 s at 72 °C and finally holding at 72 °C for 10 min. Real-time PCRs were performed  
177 with an ABI PRISM 7500 (Applied Biosystems) using SYBR Green PCR Core Reagents  
178 (Applied Biosystems) to analyze the gene expression of: i) the hepatic *vtg*, ii) *era* and  
179 *gper1*, iii) gene markers of macrophages (colony stimulating factor 1 receptor, *csflr*),  
180 IgM<sup>+</sup> B lymphocytes (immunoglobulin M heavy chain, *ighm*) and IgT<sup>+</sup> B lymphocytes  
181 (immunoglobulin T heavy chain, *ight*), iv) the pro-inflammatory cytokine interleukin 1 $\beta$ ,  
182 *illb*, and v) the anti-inflammatory cytokine *ill0*. Reaction mixtures were incubated for 10  
183 min at 95 °C, followed by 40 cycles of 15 s at 95 °C, 1 min at 60 °C, and finally 15 s at  
184 95 °C. For each mRNA, gene expression was corrected by the ribosomal protein S18 gene  
185 (*rps18*) content in each sample. The gilthead seabream specific primers used are shown  
186 in Supplementary Table 1. In all cases, each PCR was performed with triplicate samples.  
187 Less than 3% variation in the *rps18* gene expression was observed among samples.

## 188 **2.6. Statistical analysis**

189 Normal distribution of the data was analyzed with the Statgraphics Centurion XVI  
190 v16.1.15 software. Statistical differences were analyzed by Student's t tests (when only  
191 two groups were compared) or two-way ANOVA and a *post hoc* Holm-Sidak's multiple  
192 comparison test (when more than two groups were compared) using the GraphPad Prism  
193 v5.01 software.

## 194 **3. Results**

### 195 **3.1. EE<sub>2</sub> modulates the expression of hepatic vitellogenin**

196 The expression of the hepatic *vtg* gene was analyzed 76 dot and 93 dat to assess the  
197 endocrine disruption activity of dietary intake of EE<sub>2</sub>. Hepatic *vtg* mRNA levels of EE<sub>2</sub>-  
198 treated fish were much increased 76 dot compared with the untreated fish (Supplementary

199 Fig. 2A). However, no statistical differences were found at 93 dat between EE<sub>2</sub>-treated  
200 and untreated fish (Supplementary Fig. 2B).

### 201 **3.2. EE<sub>2</sub> modulates estrogen receptor expression in peritoneal leukocytes**

202 Total peritoneal leukocytes were found to express *era* and *gper1* by RT-qPCR (1 dat,  
203 1dpp) (Supplementary Fig. 3). Moreover, we observed that vaccination decreased the *era*  
204 and *gper1* transcript levels (Supplementary Fig. 3A), while EE<sub>2</sub> treatment increased the  
205 *era* and *gper1* mRNA levels in both unvaccinated and vaccinated fish (Supplementary  
206 Fig. 3A, B).

### 207 **3.3. EE<sub>2</sub> modulates leukocyte recruitment to the peritoneal cavity and peritoneal** 208 **leukocyte populations**

209 Changes in the number of peritoneal exudate leukocytes and leukocyte populations  
210 produced by EE<sub>2</sub>-treatment and vaccination were studied 1 dat (1 dpp) (Fig. 1A-F). As  
211 expected, vaccination increased the number of peritoneal leukocytes (Fig. 1A).

212 Surprisingly, dietary intake of EE<sub>2</sub> increased leukocyte recruitment in both  
213 vaccinated and unvaccinated fish (Fig. 1A). In addition, changes in leukocyte populations  
214 were also observed in peritoneal exudate 1 dpp (Fig. 1B-F). Firstly, the percentage of  
215 mast cells was substantially lower in vaccinated fish (Fig. 1B), while dietary EE<sub>2</sub> further  
216 reduced it in vaccinated fish (Fig. 1B). Secondly, vaccination promoted an increase in the  
217 percentage of acidophilic granulocytes while dietary EE<sub>2</sub> did not modify it (Fig. 1C).  
218 Thirdly, dietary EE<sub>2</sub> increased the transcript levels of genes encoding specific markers for  
219 macrophages and IgM<sup>+</sup> and IgT<sup>+</sup> B lymphocytes (*csflr*, *ighm* and *ight*, respectively) in  
220 both unvaccinated and vaccinated fish (Fig. 1D-4F). Finally, vaccination induced a  
221 reduction of the transcript levels of *ight* (Fig. 1F).

222 Ninety-three days after suspending the EE<sub>2</sub> treatment, which coincided with 1 dpb,  
223 the picture was rather similar to 1 dat (1 dpp) (Fig. 1G-1L). The most noticeable  
224 differences were that EE<sub>2</sub> increased the expression of *csflr* only in unvaccinated fish (Fig.  
225 1J), while the effect of dietary EE<sub>2</sub> in the transcript levels of *ighm* (Fig. 1K) and *ight*  
226 (Fig. 1L) were no longer observed.

227 Finally, the total leukocyte numbers were found be reduced in vaccinated and  
228 unvaccinated and EE<sub>2</sub>-treated fish at 142 days after EE<sub>2</sub> treatment, which coincided with  
229 50 dpb (data not shown).



### 230 **3.4. EE<sub>2</sub> alters the respiratory burst of peritoneal exudate leukocytes**

231 The respiratory burst was analyzed 1, 93 and 142 dat (1 dpp, 1 dpb and 50 dpb,  
232 respectively) in peritoneal exudate leukocytes (Fig. 2A-C). ROS production robustly  
233 increased 1 dat (1 dpp) and 142 dat (50 dpb) in vaccinated fish, while was slightly  
234 inhibited at 93 dat (1 dpb) (Fig. 2A-5C). In addition, dietary intake of EE<sub>2</sub> significantly  
235 induced ROS production 1 dat in unvaccinated fish (Fig. 2A), while had no effect at 93  
236 and 142 dat (Fig. 2B and 2C).

### 237 **3.5. EE<sub>2</sub> induces peritoneal *il1b* gene expression**

238 The mRNA levels of the genes encoding the pro-inflammatory IL-1 $\beta$  and the anti-  
239 inflammatory IL-10 were analyzed by RT-qPCR in total peritoneal exudate leukocytes 1,  
240 93 and 142 dat (1 dpp, 1 dpb and 50 dpb, respectively) (Fig. 2D-I). At 1 dpp (Fig. 2D)  
241 *il1b* expression was found to be higher in vaccinated fish. Strikingly, dietary intake of  
242 EE<sub>2</sub> strongly increased *il1b* transcript levels 1 (Fig. 2D) and 142 dat (Fig. 2E) in  
243 unvaccinated fish, while had no effect in vaccinated fish at any of the time analyzed (Fig.  
244 2D-F). On the other hand, *il10* expression was not altered at any of the time analyzed  
245 (Fig. 2G-I).

## 246 **4. Discussion**

247 Xeno-estrogens EDCs are able to mimic or antagonize the action of natural estrogens.  
248 That, as well as other sex steroids, have demonstrated effects beyond the reproductive  
249 context, with drastic effects along molecular to organism levels, particularly on aquatic  
250 animals (Castillo-Briceno and Kodjabachian, 2014). Despite, most fish studies on these  
251 compounds have focused on the alterations they produced on the reproductive system, it  
252 is also known that they can also alter other aspects of fish biology, including immunity  
253 (Filby et al., 2007). In this study, we analyze the effect that EE<sub>2</sub> has on the immune  
254 response of a marine hermaphrodite species, the gilthead seabream, focusing on  
255 peritoneal exudate leukocytes.

256 First of all, EE<sub>2</sub> promotes a significant increase in hepatic *vtg* gene expression in  
257 gilthead seabream, which is considered as a maker of endocrine disruption for estrogenic  
258 exposure in male fish (Sumpter and Jobling, 1995), as previously demonstrated in  
259 gilthead seabream, both in adults and juveniles, and in other fish species (Cabas et al.,  
260 2012; Kausch et al., 2008; Rodenas et al., 2016). Notably, this increase disappeared when

261 the treatment ceased, as we also previously described (Rodenas et al., 2015; Rodenas et  
262 al., 2016). Moreover, we have observed that ERs, both nuclear ERs and membrane-  
263 anchored GPER1, are expressed in peritoneal exudate leukocytes as described in mammal  
264 and in fish immune cells from head kidney (Blasko et al., 2009; Cabas et al., 2013a; Liarte  
265 et al., 2011; Straub, 2007), and that EE<sub>2</sub> increased the expression of both receptors. The  
266 present results extend these earlier observations by demonstrating that (i) EE<sub>2</sub> would exert  
267 its effect on gilthead seabream peritoneal leukocyte biology not only through the classical  
268 nuclear receptor pathway but also through the novel GPER1 and (ii) EE<sub>2</sub> would increase  
269 estrogen signaling in these cells by inducing ER. In addition, it is also interesting that  
270 vaccination decrease the expression of gene encoding ER $\alpha$  and GPER1 in peritoneal  
271 exudate leukocytes, as increases the expression of both genes in the head kidney of this  
272 species (Cabas et al., 2012; Cabas et al., 2013a). Such differences may be due to the  
273 alterations of different leukocyte populations after an immune challenge (Chaves-Pozo et  
274 al., 2005; Chaves-Pozo et al., 2004; García-Castillo et al., 2002).

275 One of the most interesting observations of this study is that dietary intake of EE<sub>2</sub>  
276 was able to promote leukocyte recruitment to the peritoneal cavity and the expression of  
277 genes encoding the *csflr*, *ighm* and *ight*, what strongly suggests that macrophages and B  
278 lymphocytes were recruited to the peritoneal cavity of unvaccinated fish, as occurred in  
279 the gonad of gilthead seabream exposed to dietary EE<sub>2</sub> (Cabas et al., 2011) and in  
280 agreement with the changes observed in leukocyte populations of Japanese sea bass blood  
281 after E<sub>2</sub> treatment (Thilagam et al., 2009). Nevertheless, once treatment had ceased, *cfsr1*  
282 gene expression remained at high levels for 3 months in EE<sub>2</sub>-unvaccinated fish, despite  
283 the number of total peritoneal exudate leukocyte returned to basal levels. Moreover, EE<sub>2</sub>  
284 increased the respiratory burst and the expression of the gene encoding IL-1 $\beta$  of peritoneal  
285 exudate leukocytes in unvaccinated fish, even 3 months after ceasing the treatment. These  
286 data are in agreement with the induction of ROS production in blood leukocytes of  
287 Japanese sea bass promoted by E<sub>2</sub> (Thilagam et al., 2009) and the induction of IL-1 $\beta$   
288 expression levels in the head kidney of gilthead seabream exposed to dietary EE<sub>2</sub> (Cabas  
289 et al., 2012). Similarly, other xeno-estrogens, such as bisphenol A, nonylphenol and  
290 different phthalates, were found to increase the ROS production by common carp  
291 phagocytic cells (Gushiken et al., 2002; Watanuki et al., 2003). The biological  
292 consequences of the long lasting effects of dietary intake of EE<sub>2</sub> shown here for the first  
293 time deserve further investigations. It is tempting to speculate that epigenetic mechanisms

294 are involved (Bhandari et al., 2015), since the endocrine disruption effect of EE<sub>2</sub> is over  
295 at this time.

296 The effect of dietary intake of EE<sub>2</sub> in vaccinated fish was also interesting, although  
297 weaker than the observed in their unvaccinated counterparts. EE<sub>2</sub> exerted its greatest  
298 effects in the total number of leukocytes and in the transcript levels of *ight 1 dpp*, as  
299 occurred in unvaccinated fish, but further decreased the percentage of mast cells.  
300 Therefore, EE<sub>2</sub> showed a synergistic effect with vaccination increasing the total number  
301 of leukocytes and reducing mast cell abundance. In contrast, other EDCs, such as PCB  
302 126, have been shown to increase the abundance of mast cells in gills and intestine of  
303 gilthead seabream (Lauriano et al., 2012).

304 To conclude, our data point to a very complex role of estrogens in fish immunity, as  
305 it is widely accepted in mammals (Straub, 2007). We found that dietary intake of EE<sub>2</sub>  
306 promotes a long lasting inflammatory response in the peritoneal cavity of unvaccinated  
307 gilthead seabream juveniles after ceasing the treatment and even though its endocrine  
308 disruption effect was over. However, the impact of dietary EE<sub>2</sub> in vaccinated fish was  
309 rather minor and transient. These results, therefore, paves the way to futures studies aimed  
310 at understanding the molecular mechanisms involved in the long lasting effect of EDCs  
311 in fish immunity.

## 312 **Acknowledgments**

313 We thank I. Fuentes for her excellent technical assistance and the “Servicio de Apoyo a  
314 la Investigación” of the University of Murcia for their assistance with cell culture, flow  
315 cytometry and real-time PCR. This work has been funded by the Fundación Séneca  
316 (Coordination Centre for Research, CARM, 19883/GERM/15 to AGA and PhD  
317 fellowship to NEGG) and the Spanish Ministerio de Economía y Competitividad  
318 (AGL2014-53167-C3-1-R, AGL2014-53167-C3-2-R) and FEDER funds.

319

## 320 **Conflict of interest**

321 The authors declare no conflict of interest.

322

## 323 **Footnotes**

324 <sup>1</sup>Abbreviations: ERs: estrogen receptors; GPER1: G protein-coupled estrogen receptor 1;  
325 E<sub>2</sub>: 17β-estradiol; EDCs: endocrine disruptor chemicals; EE<sub>2</sub>: 17α-ethynylestradiol; i.p.:  
326 intraperitoneal; PBS: phosphate buffered saline; dot: days of treatment; dat: days after  
327 treatment; dpp: days post-priming; dpb: days post-booster; mAb: monoclonal antibody;  
328 PMA: phorbolmyristate acetate; ROS: reactive oxygen species.  
329

## 330 References

- 331 Aris, A.Z., Shamsuddin, A.S., Praveena, S.M., 2014. Occurrence of 17alpha-ethynylestradiol (EE2)  
332 in the environment and effect on exposed biota: a review. *Environment International* 69, 104-  
333 119.
- 334 Baronti, C., Curini, R., D'Ascenzo, G., Di Corcia, A., Gentili, A., Samperi, R., 2000. Monitoring  
335 natural and synthetic estrogens at activated sludge sewage treatment plants and in a receiving  
336 river water. *Environmental Science and Technology* 34, 2059-5066.
- 337 Bhandari, R.K., Deem, S.L., Holliday, D.K., Jandegian, C.M., Kassotis, C.D., Nagel, S.C., Tillitt, D.E.,  
338 Vom Saal, F.S., Rosenfeld, C.S., 2015. Effects of the environmental estrogenic contaminants  
339 bisphenol A and 17alpha-ethinyl estradiol on sexual development and adult behaviors in aquatic  
340 wildlife species. *General and Comparative Endocrinology* 214, 195-219.
- 341 Blasko, E., Haskell, C.A., Leung, S., Gualtieri, G., Halks-Miller, M., Mahmoudi, M., Dennis, M.K.,  
342 Prossnitz, E.R., Karpus, W.J., Horuk, R., 2009. Beneficial role of the GPR30 agonist G-1 in an  
343 animal model of multiple sclerosis. *Journal of Neuroimmunology* 214, 67-77.
- 344 Brunsing, R.L., Owens, K.S., Prossnitz, E.R., 2013. The G protein-coupled estrogen receptor  
345 (GPER) agonist G-1 expands the regulatory T-cell population under TH17-polarizing conditions.  
346 *Journal of Immunotherapy* 36, 190-196.
- 347 Cabas, I., Chaves-Pozo, E., García-Alcázar, A., Meseguer, J., Mulero, V., García-Ayala, A., 2011.  
348 Dietary intake of 17alpha-ethynylestradiol promotes leukocytes infiltration in the gonad of the  
349 hermaphrodite gilthead seabream. *Molecular Immunology* 48, 2079-2086.
- 350 Cabas, I., Liarte, S., García-Alcázar, A., Meseguer, J., Mulero, V., García-Ayala, A., 2012. 17alpha-  
351 Ethynylestradiol alters the immune response of the teleost gilthead seabream (*Sparus aurata* L.)  
352 both in vivo and in vitro. *Developmental and Comparative Immunology* 36, 547-556.
- 353 Cabas, I., Rodenas, M.C., Abellán, E., Meseguer, J., Mulero, V., García-Ayala, A., 2013a. Estrogen  
354 signaling through the G protein-coupled estrogen receptor regulates granulocyte activation in  
355 fish. *Journal of Immunology* 191, 4628-4639.
- 356 Cabas, I., Chaves-Pozo, E., García-Alcázar, A., Meseguer, J., Mulero, V., García-Ayala, A., 2013b.  
357 The effect of 17alpha-ethynylestradiol on steroidogenesis and gonadal cytokine gene expression  
358 is related to the reproductive stage in marine hermaphrodite fish. *Marine Drugs* 11, 4973-4992.
- 359 Castillo-Briceno, P., Kodjabachian, L., 2014. *Xenopus* embryonic epidermis as a mucociliary  
360 cellular ecosystem to assess the effect of sex hormones in a non-reproductive context. *Front*  
361 *Zool* 11, 9.
- 362 Chaves-Pozo, E., Pelegrín, P., García-Castillo, J., García-Ayala, A., Mulero, V., Meseguer, J., 2004.  
363 Acidophilic granulocytes of the marine fish gilthead seabream (*Sparus aurata* L.) produce  
364 interleukin-1beta following infection with *Vibrio anguillarum*. *Cell and Tissue Research* 316, 189-  
365 195.
- 366 Chaves-Pozo, E., Muñoz, P., López-Muñoz, A., Pelegrin, P., García-Ayala, A., Mulero, V.,  
367 Meseguer, J., 2005. Early innate immune response and redistribution of inflammatory cells in  
368 the bony fish gilthead seabream experimentally infected with *Vibrio anguillarum*. *Cell and Tissue*  
369 *Research* 320, 61-68.

370 Desbrow, C., Routledge, E.J., Brighty, G.C., Sumpter, J.P., Waldock, M., 1998. Identification of  
371 estrogenic chemicals in STW effluents. 1. Chemical fractionation and *in vitro* biological  
372 screening. Environmental Science and Technology 32, 1549-1558.

373 Filby, A.L., Neuparth, T., Thorpe, K.L., Owen, R., Galloway, T.S., Tyler, C.R., 2007. Health impacts  
374 of estrogens in the environment, considering complex mixture effects. Environmental Health  
375 Perspectives 115, 1704-1710.

376 García-Castillo, J., Pelegrín, P., Mulero, V., Meseguer, J., 2002. Molecular cloning and expression  
377 analysis of tumor necrosis factor alpha from a marine fish reveal its constitutive expression and  
378 ubiquitous nature. Immunogenetics 54, 200-207.

379 Gómez-González, N.E., García-García, E., Montero, J., García-Alcázar, A., Meseguer, J., García-  
380 Ayala, A., Mulero, V., 2014. Isolation of mast cells from the peritoneal exudate of the teleost fish  
381 gilthead sea bream (*Sparus aurata* L.). Fish and Shellfish Immunology 40, 225-232.

382 Gómez-González, N.E., García-Alcázar, A., Sepulcre, M.P., Mulero, V., García-Ayala, A., 2016.  
383 Generation of a mAb, GB10, against a cell surface receptor of fish mast cell that induces cell  
384 death, in: Conference, Z. (Ed.), Mast Cell Conference 2016, Varna, Bulgaria.

385 Gushiken, Y., Watanuki, H., Sakai, M., 2002. *In vitro* effect of carp phagocytic cells by bisphenol  
386 A and nonylphenol. Fisheries Science 68, 178-183.

387 Kausch, U., Alberti, M., Haindl, S., Budczies, J., Hock, B., 2008. Biomarkers for exposure to  
388 estrogenic compounds: gene expression analysis in zebrafish (*Danio rerio*). Environmental  
389 Toxicology 23, 15-24.

390 Lauriano, E.R., Calo, M., Silvestri, G., Zaccone, D., Pergolizzi, S., Lo Cascio, P., 2012. Mast cells in  
391 the intestine and gills of the sea bream, *Sparus aurata*, exposed to a polychlorinated biphenyl,  
392 PCB 126. Acta Histochemica 114, 166-171.

393 Liarte, S., Chaves-Pozo, E., Abellan, E., Meseguer, J., Mulero, V., Canario, A.V., García-Ayala, A.,  
394 2011. Estrogen-responsive genes in macrophages of the bony fish gilthead seabream: a  
395 transcriptomic approach. Developmental and Comparative Immunology 35, 840-849.

396 Liu, X., Su, H., Zhu, P., Zhang, Y., Huang, J., Lin, H., 2009. Molecular cloning, characterization and  
397 expression pattern of androgen receptor in *Spinibarbus denticulatus*. General and Comparative  
398 Endocrinology 160, 93-101.

399 Lynn, S.G., Birge, W.J., Shepherd, B.S., 2008. Molecular characterization and sex-specific tissue  
400 expression of estrogen receptor alpha (*esr1*), estrogen receptor beta (*esr2a*) and ovarian  
401 aromatase (*cyp19a1a*) in yellow perch (*Perca flavescens*). Comparative Biochemistry and  
402 Physiology. Part B: Biochemistry and Molecular Biology 149, 126-147.

403 Meseguer, J., Esteban, M.A., Muñoz, A., López-Ruiz, A., 1993. Ultrastructure of the peritoneal  
404 exudate cells of seawater teleosts, seabream (*Sparus aurata*) and sea bass (*Dicentrarchus*  
405 *labrax*). Cell and Tissue Research 273, 301-307.

406 Milla, S., Depiereux, S., Kestemont, P., 2011. The effects of estrogenic and androgenic endocrine  
407 disruptors on the immune system of fish: a review. Ecotoxicology 20, 305-319.

408 Prossnitz, E.R., Hathaway, H.J., 2015. What have we learned about GPER function in physiology  
409 and disease from knockout mice? Journal of Steroid Biochemistry and Molecular Biology 153,  
410 114-126.

411 Quinn, B., Gagne, F., Costello, M., McKenzie, C., Wilson, J., Mothersill, C., 2004. The endocrine  
412 disrupting effect of municipal effluent on the zebra mussel (*Dreissena polymorpha*). Aquatic  
413 Toxicology 66, 279-292.

414 Revankar, C.M., Cimino, D.F., Sklar, L.A., Arterburn, J.B., Prossnitz, E.R., 2005. A transmembrane  
415 intracellular estrogen receptor mediates rapid cell signaling. Science 307, 1625-1630.

416 Rodenas, M.C., Cabas, I., Abellán, E., Meseguer, J., Mulero, V., García-Ayala, A., 2015. Tamoxifen  
417 persistently disrupts the humoral adaptive immune response of gilthead seabream (*Sparus*  
418 *aurata* L.). Developmental and Comparative Immunology 53, 283-292.

419 Rodenas, M.C., Cabas, I., García-Alcázar, A., Meseguer, J., Mulero, V., García-Ayala, A., 2016.  
420 Selective estrogen receptor modulators differentially alter the immune response of gilthead  
421 seabream juveniles. Fish and Shellfish Immunology 52, 189-197.

422 Sepulcre, M.P., Pelegrín, P., Mulero, V., Meseguer, J., 2002. Characterisation of gilthead  
423 seabream acidophilic granulocytes by a monoclonal antibody unequivocally points to their  
424 involvement in fish phagocytic response. *Cell Tissue Research* 308, 97-102.

425 Shved, N., Berishvili, G., D'Cotta, H., Baroiller, J.F., Segner, H., Eppler, E., Reinecke, M., 2007.  
426 Ethinylestradiol differentially interferes with IGF-I in liver and extrahepatic sites during  
427 development of male and female bony fish. *Journal of Endocrinology* 195, 513-523.

428 Shved, N., Berishvili, G., Hausermann, E., D'Cotta, H., Baroiller, J.F., Eppler, E., 2009. Challenge  
429 with 17alpha-ethinylestradiol (EE2) during early development persistently impairs growth,  
430 differentiation, and local expression of IGF-I and IGF-II in immune organs of tilapia. *Fish and*  
431 *Shellfish Immunology* 26, 524-530.

432 Slater, C.H., Fitzpatrick, M.S., Schreck, C.B., 1995. Characterization of an androgen receptor in  
433 salmonid lymphocytes: possible link to androgen-induced immunosuppression. *General and*  
434 *Comparative Endocrinology* 100, 218-225.

435 Straub, R.H., 2007. The complex role of estrogens in inflammation. *Endocrine Reviews* 28, 521-  
436 574.

437 Sumpter, J.P., Jobling, S., 1995. Vitellogenesis as a biomarker for estrogenic contamination of  
438 the aquatic environment. *Environmental Health Perspectives* 103 Suppl 7, 173-178.

439 Thilagam, H., Gopalakrishnan, S., Bo, J., Wang, K.J., 2009. Effect of 17beta-estradiol on the  
440 immunocompetence of Japanese sea bass (*Lateolabrax japonicus*). *Environmental Toxicology*  
441 *and Chemistry* 28, 1722-1731.

442 Thomas, P., Pang, Y., Filardo, E.J., Dong, J., 2005. Identity of an estrogen membrane receptor  
443 coupled to a G protein in human breast cancer cells. *Endocrinology* 146, 624-632.

444 Todo, T., Ikeuchi, T., Kobayashi, T., Nagahama, Y., 1999. Fish androgen receptor: cDNA cloning,  
445 steroid activation of transcription in transfected mammalian cells, and tissue mRNA levels.  
446 *Biochemical and Biophysical Research Communications* 254, 378-383.

447 Vethaak, A.D., Lahr, J., Schrap, S.M., Belfroid, A.C., Rijs, G.B., Gerritsen, A., de Boer, J., Bulder,  
448 A.S., Grinwis, G.C., Kuiper, R.V., Legler, J., Murk, T.A., Peijnenburg, W., Verhaar, H.J., de Voogt,  
449 P., 2005. An integrated assessment of estrogenic contamination and biological effects in the  
450 aquatic environment of The Netherlands. *Chemosphere* 59, 511-524.

451 Watanuki, H., Gushiken, Y., Sakai, M., 2003. *In vitro* modulation of common carp (*Cyprinus carpio*  
452 L.) phagocytic cells by Di-n-butyl phthalate and Di-2-ethylhexyl phthalate. *Aquatic Toxicology* 63,  
453 119-126.

454

455

456 **Figure legends**

457 **Figure 1.** Dietary intake of EE<sub>2</sub> promotes a peritoneal leukocyte recruitment and modulates the  
458 peritoneal populations. The total leukocyte numbers (A and G), the percentage of mast cells (MC,  
459 GB10<sup>+</sup> cells) (B and H), the percentage of acidophilic granulocytes (AG, G7<sup>+</sup> cells) (C and I), the  
460 *csflr* mRNA levels (D and J), the *ighm* mRNA levels (E and K) and the *ight* mRNA levels (F and  
461 L) at 1 and 93 dat (1 dpp and 1 dpb, respectively) in the peritoneal exudate leukocytes of fish i.p.  
462 injected with PBS (unvaccinated) or with hemocyanin plus imject alum adjuvant (vaccinated)  
463 previously exposed to dietary EE<sub>2</sub> (EE<sub>2</sub>, solid bars) or not (C, open bars). dat: days after treatment;  
464 dpp: days post-priming; dpb: days post-booster. Data are mean ± SEM of 6 individuals. The  
465 asterisks denote statistically significant differences between groups according to two-way  
466 ANOVA and Holm-Sidak's multiple comparison tests. ns, not significant; \*p < 0.05;  
467 \*\*p<0.01 and \*\*\*\*p<0.0001.

468 **Figure 2.** Dietary intake of EE<sub>2</sub> modulates the respiratory burst activity and induces *illb* gene  
469 expression of peritoneal exudate leukocytes. ROS production (A-C) triggered by PMA was  
470 measured as maximum values of luminol- dependent luminescence at 1 (A), 93 (B) and 142 (C)  
471 dat (1 dpp, 1 dpb and 50 dpb, respectively) generated by total peritoneal exudate leukocytes of  
472 fish i.p. injected with PBS (unvaccinated) or with hemocyanin plus imject alum adjuvant  
473 (vaccinated) previously exposed to dietary EE<sub>2</sub> (EE<sub>2</sub>, solid bars) or not (C, open bars). The mRNA  
474 levels of the genes encoding IL-1β and IL-10 (D-I) was analyzed by RT-qPCR at 1 (A and D), 93  
475 (B and E) and 142 (C and F) dat (1 dpp, 1 dpb and 50 dpb, respectively) in total peritoneal exudate  
476 leukocytes of fish i.p. injected with PBS (unvaccinated) or with hemocyanin plus imject alum  
477 adjuvant (vaccinated) previously exposed to dietary EE<sub>2</sub> (EE<sub>2</sub>, solid bars) or not (C, open  
478 bars).dat: days after treatment; dpp: days post-priming; dpb: days post-booster. Data are mean ±  
479 SEM of 6 individuals. The asterisks denote statistically significant differences between groups  
480 according to two-way ANOVA and Holm-Sidak's multiple comparison tests. ns, not  
481 significant; \*p < 0.05; \*\*p<0.01 and \*\*\*\*p<0.0001.

482 **Supplementary Figure 1.** Schematic drawing of the experimental design. Animals were exposed  
483 to 0 (untreated fish) and 5 (treated fish) µg EE<sub>2</sub>/g food for 76 days (days of treatment, dot), after  
484 which they were fed with commercial food for a further 142 days. Fish were i.p. injected with  
485 PBS (unvaccinated fish) or with hemocyanin plus imject alum adjuvant (vaccinated fish) at 76  
486 dot (priming) and 92 days after treatment (dat) (booster). Samples were collected 1, 93 and 142  
487 dat (1 day post-priming, dpp, 1 day post-booster, dpb, and 50 dpb, respectively).

488 **Supplementary Figure 2.** Dietary intake of EE<sub>2</sub> increases the expression of the gene encoding  
489 vitellogenin (*vtg*) in gilthead seabream liver. RT-qPCR analysis of hepatic *vtg* mRNA levels in  
490 untreated (C, open bars) and dietary EE<sub>2</sub>-treated fish (EE<sub>2</sub>, solid bars) at 76 days of treatment (dot)  
491 (A) and 93 days after treatment (dat) (B). Data are mean ± SEM of 6 individuals. The asterisks  
492 denote statistically significant differences between groups according to Student's t-test. ns, not  
493 significant and \*\*\*p<0.001.

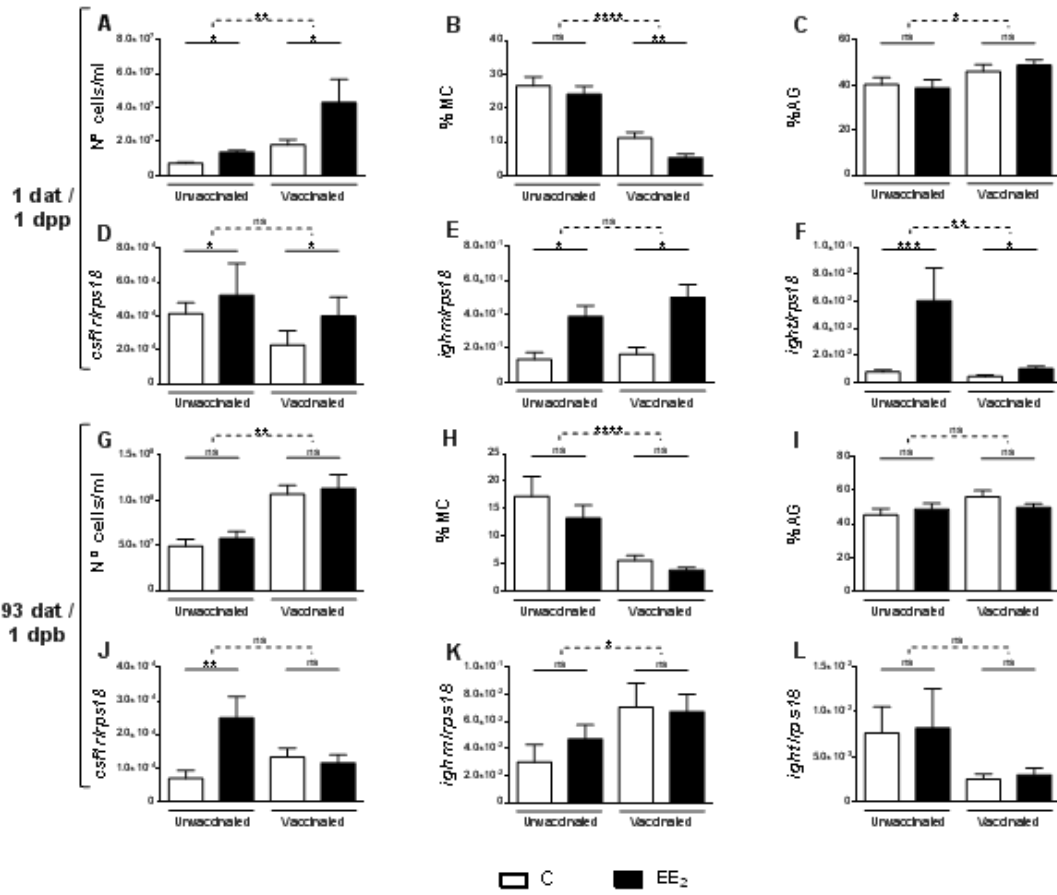
494 **Supplementary Figure 3.** Dietary intake of EE<sub>2</sub> alters the mRNA expression levels of the  
495 estrogen receptors *era* and *gper1* in peritoneal exudate. RT-qPCR analysis of *era* (A) and *gper1*  
496 (B) mRNA levels at 1 dat (1dpp) in the peritoneal exudate leukocytes of fish i.p. injected with  
497 PBS (unvaccinated) or hemocyanin plus imject alum adjuvant (vaccinated) previously exposed  
498 (EE<sub>2</sub>, solid bars) or not (C, open bars) to dietary EE<sub>2</sub>. dat: days after treatment; dpp: days post-  
499 priming; dpb: days post-booster. Data are mean ± SEM of 6 individuals. The asterisks denote  
500 statistically significant differences between groups according to two-way ANOVA and Holm-  
501 Sidak's multiple comparison tests. \*p < 0.05; \*\*p<0.01 and \*\*\*p<0.001.

502

503

504 **Figure 1**

505



506

507

508

509

510

511

512

513

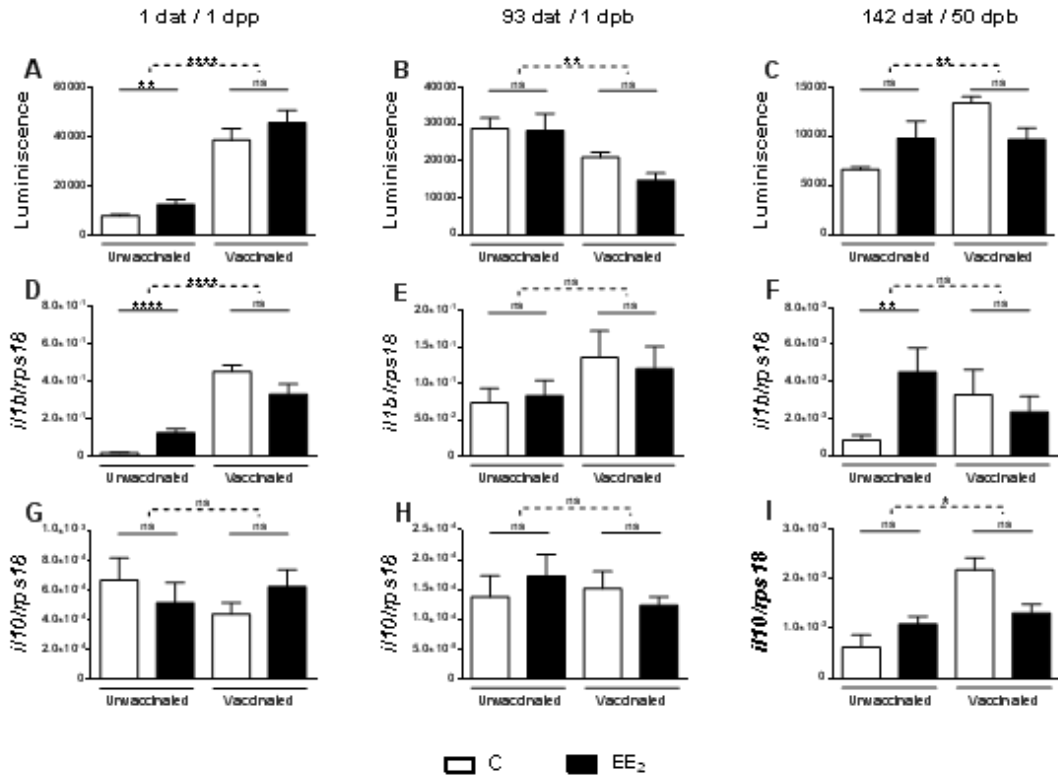
514

515

516



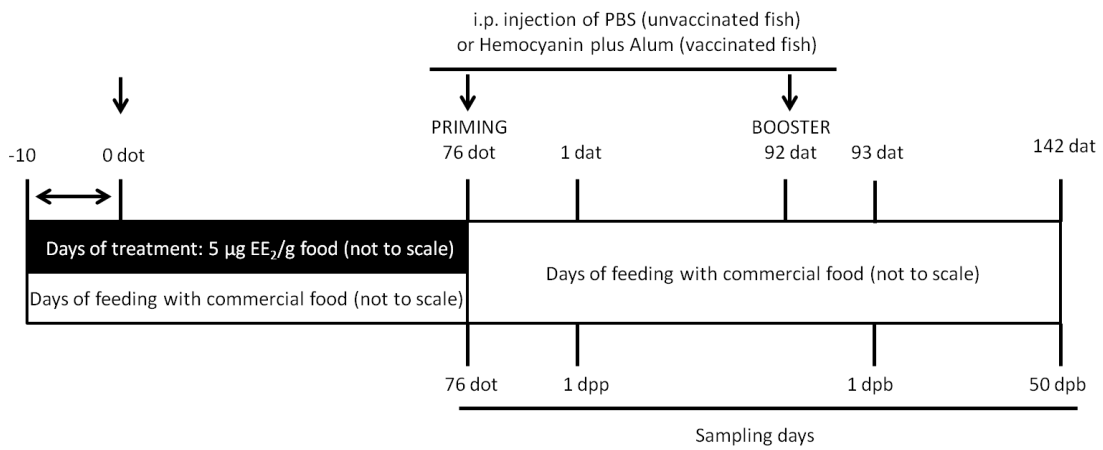
517 **Figure 2**



518

519 **Supplementary Figure 4**

520



521

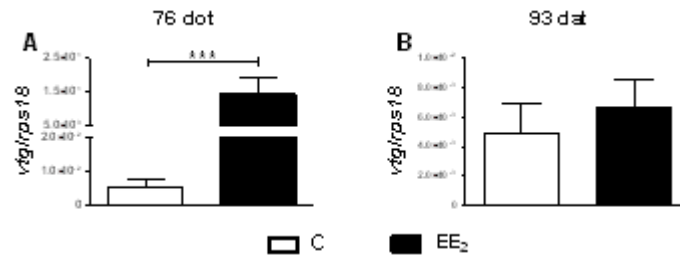
522

523

524

525

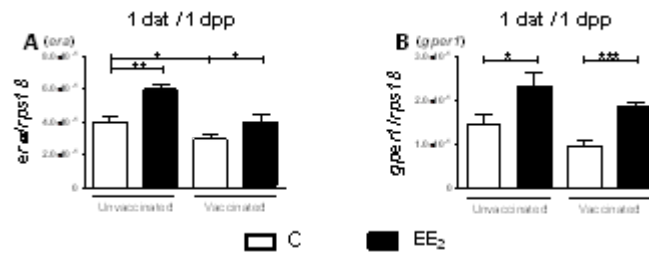
526 **Supplementary Figure 2**



527

528

529 **Supplementary Figure 3**



530